Different Patterns Of Nutrient Cycling In Contiguous Phytophysiognomies Of Atlantic Forest, Brazil

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ABSTRACT
The present study aimed to evaluate fine aboveground litterfall, deposition of nitrogen and phosphorus from total litterfall and leaf litter decomposition in areas of tall forest (Mata Alta) and low forest on sandy soils (Mussununga) in southeastern Brazil. Fine litterfall was collected monthly for two years (from June/2007 to May/2009) in 10 conical collectors (0.25 m²) in each phytophysiognomy (1 ha plots). The material was subsequently separated into leaves, branches, flowers and fruits, and unidentified material. Leaf decomposition rates were evaluated using 15 litterbags, three of which were collected every 30 days. Higher litterfall occurred in both phytophysiognomies during the rainy season, with leaves predominating. The lowest litterfall, nutrient input, and leaf litter decomposition values appeared to be nutrient conservation strategies, contributing to the ecological functioning of the Mussununga where soil fertility was lower than in the Mata Alta.

Keywords: decomposition, ecological functioning, litterfall, tropical forest.
1. INTRODUCTION

Litterfall and decomposition play critical roles in the productivity of forest systems and help promote tropical forest growth through efficient nutrient recycling (Pimenta et al., 2011; Marafiga et al., 2012). However, the naturally low fertility of most tropical forest soils can limit litterfall and nutrient input (Kaspari et al., 2008). Estimates of litterfall and decomposition rates, and the nutrient content of litter, are fundamental to our understanding of the biogeochemical functioning of those ecosystems therefore, especially given predicted climatic alterations (Parsons et al., 2014).

On a global scale, climate exerts a significant influence on the ecological processes of litterfall (Zhang et al., 2008) and decomposition (Aerts, 1997). At a local scale, microclimatic conditions, such as soil fertility (Parsons et al., 2014), soil moisture, and temperature, which depend on the structure and diversity of the arboreal community, control litterfall and decomposition (Menezes et al., 2010; Bianchin et al., 2016). The diversity of tree species strongly influences litter decomposition at smaller spatial scales (Scheer et al., 2011; Bonanomi et al., 2013) due to variations in litter chemical quality, which is related to nutrient concentrations and C:N ratios (Zhang et al., 2014).

There is information regarding fine litterfall, chemical content, and/or litter decomposition in different Atlantic Forest physiognomies (Moraes et al., 1999; Menezes et al., 2010; Pimenta et al., 2011; Scheer et al., 2011; Marafiga et al., 2012; Pereira et al., 2012; Freire et al., 2014; Sloboda et al., 2017). Fine litterfall is composed of leaves, branches ≤ 2 cm in diameter, reproductive material, and unidentified material < 2 mm (Vitousek, 1984). However, few studies have focused on nutrient cycling in contiguous vegetation formations in this biome that experience the same climatic regimes but demonstrate abrupt variations in terms of vegetation structure, floristic composition, and soil fertility.

The present study aimed to evaluate fine aboveground litterfall, nutrient input, and leaf litter decomposition rates in two contiguous physiognomies of Atlantic Forest that differ strongly in terms of their floristics and soil fertility, seeking to better understand the spatial variations that can occur in those key processes of ecosystem functioning.

We tested the hypothesis that fine aboveground litterfall, nutrient input, and leaf litter decomposition rates are lower in a low forest area on sandy soils (Mussununga) due to the lower natural soil fertility and lower forest structure in comparison to a contiguous tall forest area (Mata Alta), in southeastern Brazil.

2. MATERIAL AND METHODS

2.1. Study area

The present study was undertaken in the Vale Natural Reserve which occupies an area of approximately 22,000 ha, in the municipality of Linhares (19° 06’ - 19° 18’ S; 39° 45’- 40° 19’ W), in Espírito Santo State, Brazil. The regional climate is tropical hot and humid (type Aw), with a rainy austral summer season and dry austral winter (Alvares et al., 2013). The total annual rainfall and mean temperature during the first study year (year 1: June/2007 to May/2008) were 975.8 mm and 24.9 °C, respectively (Figure 1);
and 1,337.4 mm and 25.2 °C, respectively during the second year (year 2: June/2008 to May/2009). As such, the accumulated rainfall in year 1 was approximately 362 mm less than year 2. The climatic data were obtained from the weather station at Linhares maintained by the Capixaba Institute for Research, Technical Assistance and Rural Extension (Incaper), which is located approximately 27 km away from the study area. According to the historical series, the total annual rainfall and mean temperature for the areas were 1,159.0 mm and 24.4 °C, respectively (Clima Tempo, 2019).

The Reserve is dominated by Seasonal Semi-Deciduous Forest, also known as “Tabuleiro” (upland) forest, although other forest physiognomies such as Mata Alta (Tall Forest) and Mussununga are also present. We selected two contiguous one-hectare areas in these latter phytophysiognomies. The areas of Mata Alta present high tree species richness (271 species distributed among 55 families ha⁻¹), which can reach up to 40 m in height and form a closed canopy (Peixoto et al., 2008). Mussununga forests occur in patches within the Mata Alta, with trees presenting heights of between 7 and 10 m (rarely reaching 20 m) that form a discontinuous canopy with lower species richness than the Mata Alta (79 species distributed among 29 families ha⁻¹) (Simonelli et al., 2008).

Mussununga shows higher floristic similarity with nearby areas of restinga (nearshore, sandy-soil vegetation) than with Mata Alta (Giaretta et al., 2013). Mata Alta forests occur on clayey or clayey-sandy soils, classified as Yellow Argisol (Ultisol), while Mussununga vegetation occurs on sandy Spodosol (Garay et al., 1995). According to the abovementioned authors, the topsoil in the Mata Alta demonstrates higher fertility when compared to the Mussununga area (Table 1).

2.2. Fine litterfall

In each area, the fine aboveground litterfall was sampled using 10 conical collectors (0.25 m⁻² in diameter) made of 1 mm nylon mesh, at 1.30 m above ground level. The collectors were allocated at a distance of 50 m from each other, distributed in three parallel delineated lines, spaced at a distance of 150 m from each other, in the center of each area. This systematic arrangement of collectors aimed to avoid the edge effect on litterfall. This material was collected monthly during years 1 and 2 and was taken to the laboratory for separation into fractions: leaf, branch (≤ 2 cm in diameter), flower and fruit, and unidentified material < 2 mm fractions. Afterwards, the material was dried in a forced air circulation oven (70 °C, 72 h).

We calculated the daily fine litterfall (g m⁻² day⁻¹) using the equation:

\[
DFL = \left(\frac{\text{mean of litterfall}}{\text{CA}}\right) / \text{IDC}
\]

where: DFL = daily fine litterfall, CA = collector area ( = 0.25 m²), and IDC = interval, in days, between two successive collections (= 30 days).

The mean annual fine litterfall (Mg ha⁻¹ year⁻¹) was then estimated using the equation (Lopes et al., 2002):

\[
AFL = \frac{\text{ML} \times 10,000}{\text{CA}}
\]

where: AFL = mean annual fine litterfall, MFL = mean monthly fine litterfall (Mg ha⁻¹ month⁻¹), and CA = collector area.

2.3. Nutrients in total fine litterfall

We pooled the total monthly litterfall fractions obtained in year 1 to form a composite sample per collector, in each area. Then, three out of 10 collectors were randomly selected per month in each area to obtain three samples composed of total litterfall to determine nitrogen (N) and phosphorus (P) concentrations (Tedesco et al., 1995). This procedure aimed to optimize chemical analysis. Afterwards, we estimated the nutrient content in total litterfall based on the product between the total litterfall dry mass deposited monthly in the collectors and its nutrient

<table>
<thead>
<tr>
<th>Phytophysiognomy</th>
<th>Clay</th>
<th>Silt</th>
<th>Sand</th>
<th>C</th>
<th>N</th>
<th>C/N</th>
<th>pH</th>
<th>P (mg kg⁻¹)</th>
<th>K⁺</th>
<th>Ca²⁺</th>
<th>Mg²⁺</th>
<th>Na⁺</th>
<th>H⁺+Al³⁺</th>
<th>SB (cmol kg⁻¹)</th>
<th>CEC</th>
</tr>
</thead>
<tbody>
<tr>
<td>Mata Alta</td>
<td>7</td>
<td>5</td>
<td>88</td>
<td>0.86</td>
<td>0.07</td>
<td>12</td>
<td>5.6</td>
<td>2.80</td>
<td>0.07</td>
<td>1.73</td>
<td>0.44</td>
<td>0.05</td>
<td>2.61</td>
<td>2.29</td>
<td>4.9</td>
</tr>
<tr>
<td>Mussununga</td>
<td>1</td>
<td>5</td>
<td>94</td>
<td>1.16</td>
<td>0.07</td>
<td>17</td>
<td>4.6</td>
<td>1.15</td>
<td>0.05</td>
<td>0.38</td>
<td>0.50</td>
<td>0.04</td>
<td>6.03</td>
<td>0.97</td>
<td>7.0</td>
</tr>
</tbody>
</table>

SB = sum of bases; CEC = total cation exchange capacity.
content, in each area. We presented the mean annual values of both nutrient concentration and content in the total litterfall.

2.4. Leaf litter decomposition

Leaf litter decomposition was evaluated using 15 litterbags placed on the ground near the conical collectors in each physiognomy. The bags (25 × 25 cm) were made from 4 mm nylon mesh (Menezes et al., 2010). Approximately 10 g of leaf litter obtained from the conical collectors was placed in each sack. Three litterbags were collected every 30 days for 150 days. The remaining materials in the litterbags at each collection time were oven dried (65 ºC, 48 h) and weighed, resulting in measures of mass loss by subtracting the weight at each collection period from the initial mass (10 g).

The decomposition constant (g g⁻¹ day⁻¹) was estimated using the equation of decreasing mass (Thomas & Asakawa, 1993):

\[ X_t = X_0 \times e^{-kt} \]  

(3)

where: \( X_t \) = dry mass of the material remaining after \( t \) days, \( X_0 \) = the dry mass of the material originally placed in the sacks at \( t_0 \), and \( k \) = decomposition constant.

The half-life of the material \( T_{1/2} \), corresponding to the time required for the decomposition of 50% of the original mass, was calculated using the \( k \) value obtained from the following equation (Rezende et al., 1999):

\[ T_{1/2} = \left[ \frac{ln (2)}{k} \right] \]  

(4)

where: \( T_{1/2} \) = half-life of the material, and \( k \) = decomposition constant.

2.5. Statistical analysis

The data for fine litterfall (total and fractions) were submitted to the Kolmogorov-Smirnov test for normality. We evaluated the effects of the physiognomy (Mata Alta × Mussununga) and the month of collection within each year (year 1 × year 2) on fine litterfall by ANOVA Repeated Measures Design. We considered the values of fine litterfall (total and fractions) obtained from the collectors per month of the year in question as dependent variables and the collection dates (the total number of collection months in the year) as “within effects”; the phytophysiognomy was considered a “between effects” predictor.

When we observed a significant effect from the phytophysiognomy on annual fine litterfall (total and fractions), we proceeded with the analysis of variance (One-Way ANOVA), using the Levene test. We compared the means using the Student t test when the variances demonstrated homoscedasticity, or the Mann-Whitney nonparametric test when that premise was not respected. We also verified the effect of phytophysiognomy on the nutrients in the total fine litterfall as described above. The influence of climatic factors (monthly rainfall and mean temperatures) on litterfall (total and leaves) was evaluated using the Spearman correlation test. We considered the climatic data obtained for years 1 and 2 and the historical series. All of those analyses were run on STATISTICA version 8.0 software, considering \( p < 0.05 \).

In relation to the decomposition experiment, we obtained the mass loss curve, as well as the exponential model that resulted in the \( k \) value and the \( T_{1/2} \), using analysis of variance (F test, \( P < 0.05 \)) of the rates of decomposition, using SigmaPlot software for Windows version 12.0.

3. RESULTS

3.1. Fine litterfall

We observed significant interactions between phytophysiognomy and the collection months in year 1 for total litterfall (\( F = 3.1286; P = 0.000649 \)), leaves (\( F = 5.3474; P = 0.000000 \)), and branches (\( F = 2.68976; P = 0.003026 \)); and in the means between the years for total litterfall (\( F = 2.4865; P = 0.006074 \)), leaves (\( F = 4.1297; P = 0.000018 \)), and branches (\( F = 3.51886; P = 0.000160 \)). There was also significant effect of the collection month in year 2 for total litterfall (\( F = 10.1567; P = 0.000000 \)), leaves (\( F = 5.3474; P = 0.000000 \)), and branches (\( F = 2.68976; P = 0.003026 \)); and in the means between the years for total litterfall (\( F = 18.0145; P = 0.000000 \)).

We also observed a significant effect of the collection month in year 2 for total litterfall (\( F = 10.1567; P = 0.000000 \)), leaves (\( F = 27.2023; P = 0.000000 \)), branches (\( F = 3.16577; P = 0.000568 \)), and in the mean between the years for total litterfall (\( F = 18.0145; P = 0.000000 \)).
P = 0.000000, leaves (F = 37.1201; P = 0.000000), branches (F = 3.4786; P = 0.000185), and unidentified material (F = 2.5746; P = 0.004499).

Total litterfall was significantly higher in the Mata Alta forest (years 1 and 2, mean between the years), when compared to the Mussununga (Table 2). We also observed higher values for branch (year 1), and flower and fruit (year 2 and means between the years) deposition in the Mata Alta. There were no significant differences between the phytophysiognomies in terms of the leaf and unidentified material fractions during those years or in the means between the years.

Leaves represented more than 50% of the total annual litterfall in the phytophysiognomies during both year 1 (Mata Alta: 65%; Mussununga: 68%) and year 2 (Mata Alta: 56%; Mussununga: 62%) (Table 2). Considering the means for those two years, leaves contributed 62 and 65% of the total annual litter deposited in the Mata Alta and Mussununga forests, respectively.

A significant effect was observed in terms of the collection year on litterfall in both the Mata Alta and the Mussununga forests. The Mata Alta phytophysiognomy demonstrated significantly higher leaf and total litterfall deposition during year 1 when compared to year 2 (Table 2). The same tendency was observed in the Mussununga forest, with significantly higher leaf, flower and fruit, and total litterfall deposition in year 1 than in year 2.

In general, we observed the highest leaf biomass and total litterfall in the months with higher rainfall. During year 1, the highest leaf and total litterfall deposition in the Mata Alta forest occurred between November/2007 and January/2008, while the peak of leaf and total litterfall deposition in the Mussununga forest occurred in December/2007 (Figure 2). The highest leaf and total litterfall during year 2 were observed between January/2008 and February/2008 in both phytophysiognomies.

Considering the climatological data from the study period, we observed a significant positive correlation between mean accumulated monthly precipitation and leaf deposition, in the Mata Alta (years 1 and 2) and Mussununga (year 2) (Table 3). A significant positive correlation was also observed between precipitation and total litterfall, in the Mussununga (year 2). There was also a significant positive correlation between the mean monthly temperatures and leaf deposition in the Mata Alta (year 1) and Mussununga (year 2), and total litterfall in the Mussununga (year 2).

In relation to the historical series, we observed a significant positive correlation between precipitation and leaf deposition in the Mata Alta (both years 1 and 2) and in the Mussununga (year 2), and between precipitation and total litterfall in the Mata Alta (year 1) and Mussununga (year 2) (Table 3). The same significant positive correlation between mean temperature and leaf deposition and total litterfall occurred in both Mata Alta and Mussununga. There was also a

Table 2. Annual litterfall (total and fractions) during year 1 (June/2007 to May/2008) and year 2 (June/2008 to May/2009) in the Mata Alta and Mussununga phytophysiognomies within the upland (Tabuleiro) forests of the Vale Natural Reserve, in Linhares, Espírito Santo State, Brazil*.

<table>
<thead>
<tr>
<th>Phytophysiognomy</th>
<th>Leaf</th>
<th>Branch</th>
<th>Flower and fruit</th>
<th>Unidentified material</th>
<th>Total</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td><strong>Year 1</strong></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Mata Alta</td>
<td>8.96 Aa</td>
<td>3.47 Aa(2.20)</td>
<td>1.10 Aa (0.56)</td>
<td>0.97 Aa (0.30)</td>
<td>14.50 Aa (3.21)</td>
</tr>
<tr>
<td>Mussununga</td>
<td>8.20 Aa</td>
<td>1.58 Ba (1.37)</td>
<td>0.79 Aa (0.53)</td>
<td>0.83 Aa (0.54)</td>
<td>11.40 Ba (3.07)</td>
</tr>
<tr>
<td><strong>Year 2</strong></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Mata Alta</td>
<td>5.43 Ab</td>
<td>1.69 Aa (0.49)</td>
<td>1.32 Aa (0.96)</td>
<td>1.26 Aa (0.72)</td>
<td>7.91 Ab (1.73)</td>
</tr>
<tr>
<td>Mussununga</td>
<td>4.28 Ab</td>
<td>1.49 Aa (1.59)</td>
<td>0.36 Bb (0.32)</td>
<td>0.83 Aa (0.34)</td>
<td>6.96 Bb (2.98)</td>
</tr>
<tr>
<td><strong>Mean</strong></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Mata Alta</td>
<td>7.20 A</td>
<td>2.58 A (1.14)</td>
<td>1.21 A (0.61)</td>
<td>1.11 A (0.36)</td>
<td>12.10 A (1.71)</td>
</tr>
<tr>
<td>Mussununga</td>
<td>6.24 A</td>
<td>1.54 A (1.33)</td>
<td>0.57 B (0.34)</td>
<td>0.83 A (0.40)</td>
<td>9.18 B (2.74)</td>
</tr>
</tbody>
</table>

*The values indicate the means obtained from 10 replicates, followed by the standard deviation (in parentheses). Values followed by different uppercase letters within the same litter fraction and the same year indicate significant differences between the phytophysiognomies, as determined by the parametric Student t test or the nonparametric Mann-Whitney test (P < 0.05). Values followed by different lowercase letters within the same litter fraction and the same phytophysiognomy indicate significant differences between those years by the parametric Student t test or the nonparametric Mann-Whitney U test (P < 0.05).
significant positive correlation between the data from the historical series and the precipitation data from year 1 (0.63, P < 0.05) and year 2 (0.83, P < 0.05), and temperature from historical series and temperature from both years 1 and 2 (0.94, P < 0.05).

### 3.2. Nutrients in total fine litterfall

The mean values for nitrogen concentration and content were both higher in the total fine litterfall in the Mata Alta area, compared to the Mussununga area (Table 4). On the other hand, there were no significant differences between the areas in terms of phosphorus concentration and content in the total fine litterfall.

### 3.3. Leaf litter decomposition

We observed contrasting patterns of leaf litter decomposition in the two phytophysiognomies studied. There was a rapid loss of leaf litter mass during the first 60 days in the Mata Alta, with only approximately 82% of the original mass remaining after this period (Figure 3).
In the Mussununga forest, however, approximately 89% of the original leaf litter mass was still present after the first 60 days. After 150 days, the remaining leaf litter material in the Mata Alta forest was only approximately 59% of the original mass, while 72% of the original leaf litter mass remained in the Mussununga forest (Figure 3). The value of the decomposition constant \( k \) in the Mata Alta was 0.0036 g g\(^{-1}\) day\(^{-1}\), that is, 2 times higher than that calculated for the Mussununga forest (0.0019 g g\(^{-1}\) day\(^{-1}\)). As such, the half-life \( T_{1/2} \) of leaf litter in the Mussununga forest was 365 days, approximately twice that observed in the Mata Alta (193 days). The leaf mass decrease equations in the Mata Alta and Mussununga forests showed significant \( (P < 0.05) \) and high values for correlation coefficients \( (R > 0.80) \).

### Table 4. Concentration and content of nitrogen (N) and phosphorus (P) in total fine litterfall, in the Mata Alta and Mussununga phytophysiognomies within the upland (Tabuleiro) forests of the Vale Natural Reserve, in Linhares, Espírito Santo State, Brazil*

<table>
<thead>
<tr>
<th>Phytophysiognomy</th>
<th>N (g kg(^{-1}))</th>
<th>N (kg ha(^{-1}) year(^{-1}))</th>
<th>P (g kg(^{-1}))</th>
<th>P (kg ha(^{-1}) year(^{-1}))</th>
</tr>
</thead>
<tbody>
<tr>
<td>Mata Alta</td>
<td>14.71 A</td>
<td>212.05 A</td>
<td>0.95 A</td>
<td>13.84 A</td>
</tr>
<tr>
<td>Mussununga</td>
<td>9.54 B</td>
<td>106.68 B</td>
<td>0.91 A</td>
<td>10.20 A</td>
</tr>
</tbody>
</table>

*The values indicate the means obtained from three replicates. Values followed by different letters indicate significant differences between the phytophysiognomies, as determined by the parametric Student t test or the nonparametric Mann-Whitney test \( (P < 0.05) \).”

**Figure 3.** Decomposition of the leaf litter fraction (g) during the study period (150 days) in the Mata Alta and Mussununga phytophysiognomies within the upland (Tabuleiro) forests of the Vale Natural Reserve, in Linhares, Espírito Santo State, Brazil.

### 4. DISCUSSION

#### 4.1. Fine litterfall

The higher total annual litterfall in the Mata Alta was a result of the more robust structure of that phyt Physiognomy, which, in turn, led to the differences between the areas in terms of primary productivity (Capelless et al., 2016) related to their contrasting edaphic characteristics. The soil in the Mussununga area is sandy and nutrient-poor when compared to the Mata Alta soil (Garay et al., 1995) due to cation leaching. The primary productivities of tropical forests are lower in soils presenting lower phosphorus, potassium, and calcium availability (Moraes et al., 1999; Chave et al., 2010; Baribault et al., 2012). The annual fine litterfall in Mata Alta and Mussununga forests were both higher
when compared to the estimated means for other Seasonal Forest areas and for the different Atlantic Forest phytophysiognomies (Table 5).

As mentioned earlier, the Mussununga area presents high floristic similarity with sandy, nearshore restinga vegetation (Giaretta et al., 2013). In spite of this, and regardless of both ecosystems growing on nutrient-poor sandy soils, the total litterfall of the Mussununga area was 64% higher (mean of the study years) than the mean value for restinga forests (Table 4). This finding demonstrates, once again, the influence of arboreal community structure on litterfall. In fact, the canopy in the Mussununga area is higher (7-15 m) when compared with non-flooding areas of restinga (4-7 m) on Ilha do Cardoso and Ilha do Mel (Moraes et al., 1999). On the other hand, the mean annual total litterfall in the Mussununga forest was very close to a dune forest in the Restinga da Marambaia (6.8 Mg ha⁻¹) whose canopy height (8-15 m) (Camara et al., 2018a) was similar to the Mussununga forest.

The higher relative contribution of the leaf fraction to the annual total litterfall in both areas represents a pattern commonly reported in the literature (Table 4). The sandy textures of Mussununga soils apparently also influenced the higher contribution of leaves to total litterfall in relation to the Mata Alta forest. Soils with higher sand content and lower clay content retain less water (Pastor & Post, 1986) and allow increased litter deposition from plants (Souza et al., 2019). Thus, the arboreal community in the Mussununga phytophysiognomy likely demonstrates higher leaf abscission than the Mata Alta area to reduce water loss through evapotranspiration.

Litterfall occurred continuously during the present study, although we observed higher deposition during the rainy season, probably reflecting an increase in leaf renewal, which occurs simultaneously with growth (due to the higher availability of water in the soil). Additionally, high winds are very common during the rainy season, as well as torrential downpours, whose

Table 5. Annual total litterfall (Mg ha⁻¹ year⁻¹) and the relative contribution of the leaf fraction (%) in different Atlantic Forest phytophysiognomies (SSDF: Seasonal Semi-Deciduous Forest; SDF: Seasonal Deciduous Forest; NDF: Non-Deciduous Forest; RE: Restinga), Brazil.

<table>
<thead>
<tr>
<th>Phytophysiognomy</th>
<th>Site</th>
<th>Total litterfall</th>
<th>Leaves</th>
<th>Reference</th>
</tr>
</thead>
<tbody>
<tr>
<td>SSDF</td>
<td>Linhares, ES</td>
<td>12.1 (MA)¹</td>
<td>60⁰</td>
<td>This study</td>
</tr>
<tr>
<td></td>
<td>Interior of PR</td>
<td>9.2 (MU)³</td>
<td>68¹</td>
<td>Pimenta et al. (2011)</td>
</tr>
<tr>
<td></td>
<td>Pinheiral, RJ</td>
<td>8.2</td>
<td>79⁰</td>
<td>Menezes et al. (2010)</td>
</tr>
<tr>
<td></td>
<td>Pinheiral, RJ</td>
<td>6.6 (FSEI)²</td>
<td>84 (FSEI)²</td>
<td>Menezes et al. (2010)</td>
</tr>
<tr>
<td></td>
<td>Pinheiral, RJ</td>
<td>7.4 (FSEM)²</td>
<td>75 (FSEM)²</td>
<td>Menezes et al. (2010)</td>
</tr>
<tr>
<td></td>
<td>Pinheiral, RJ</td>
<td>11.0 (FSEA)²</td>
<td>53 (FSEA)²</td>
<td>Menezes et al. (2010)</td>
</tr>
<tr>
<td>SDF</td>
<td>Itaara, RS</td>
<td>5.9</td>
<td>63</td>
<td>Marafiga et al. (2012)</td>
</tr>
<tr>
<td>Mean</td>
<td></td>
<td>8.6</td>
<td>69</td>
<td></td>
</tr>
<tr>
<td>NDF</td>
<td>Rio de Janeiro, RJ</td>
<td>5.6</td>
<td>61</td>
<td>Freire et al. (2014)</td>
</tr>
<tr>
<td></td>
<td>Ilha do Cardoso, SP</td>
<td>6.3</td>
<td>70</td>
<td>Moraes et al. (1999)</td>
</tr>
<tr>
<td></td>
<td>Guaraqueçaba, PR</td>
<td>6.4 (Site 1)³</td>
<td>69 (Site 1)³</td>
<td>Scheer et al. (2011)</td>
</tr>
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<td>Antonina, PR</td>
<td>7.7 (FS-1)⁴</td>
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<td>67 (FS-2)⁴</td>
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<td>66 (FS-3)⁴</td>
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¹MA and MU: Mata Alta and Mussununga, respectively. ²FSEI, FSEM, and FSEA: secondary forest initial, medium, and advanced stage. ³Site 1 and Site 2: most advanced and least advanced secondary successional stage, respectively. ⁴FS-1, FS-2, and FS-3: Less evolved, intermediate and more evolved secondary forests, respectively. ⁵Periodically flooded forest.
mechanical impacts tend to produce high litterfall (Camara et al., 2018b).

4.2. Nutrients in total fine litterfall

The highest nutrient input from total fine litterfall to the soil was proportional to the higher amount of litterfall in the Mata Alta area. This pattern is commonly observed in the literature (Pimenta et al., 2011; Sloboda et al., 2017). The low natural fertility of the soil in the Mussununga forest also influenced the lower concentrations of nutrients in its litterfall (Moraes et al., 1999). The nitrogen concentration in total fine litterfall in the Mussununga area was much lower and approached the value observed in the restinga located in Cardoso Island, SP (Table 6).

On the other hand, nitrogen concentration in total fine litterfall in the area of Mata Alta approached the estimated mean value for different areas of Semi-Deciduous and Deciduous Seasonal Forest, and Non-Deciduous Forest (Table 5). Regarding the phosphorus concentration, both areas of the present study showed high values compared to the different Atlantic Forest phytophysiognomies.

4.3. Leaf litter decomposition

Litter decomposition was faster in the first trimester, with a subsequent slowing in both phytophysiognomies. The initial phase of fast decomposition, which is due to the high content of rapidly decomposable components, is followed by a lower rate of mass loss due to remaining, more resistant structures (Matos et al., 2011). There was an adequate adjustment of the data obtained from the decomposition experiment to the model of leaf litter mass decay.

Results obtained in the different areas of Atlantic Forest and restingas corroborated the relatively rapid rate of decomposition observed in the Mata Alta (Table 7). On the other hand, leaf litter decomposition in the Mussununga area was considered slow, with the decomposition rate being close to the value observed in a periodically flooded area of restinga forest (Pereira et al., 2012). The low nutrient quality of litter (Castanho & Oliveira, 2008; Pereira et al., 2012) can retard its breakdown and slow its decomposition in some forest ecosystems (Zhang et al., 2008) such as in restingas (Camara et al., 2018a). In fact, the same pattern occurred in the Mussununga area, where lower nitrogen concentrations were observed, which led to lower rates of leaf litter decomposition compared to the Mata Alta area. Vegetation growing on sandy soils presented lower nitrogen and phosphorus concentrations in the leaves than the vegetation established on clayey soils (Thompson et al., 1992).

According to Garay et al. (1995), the litter standing stock on the soil surface in the Mussununga area (512 kg ha⁻¹) was almost four times greater than in the Mata Alta area (141 kg ha⁻¹), resulting in greater organic material stock in the soil of the former

<table>
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<th>Phytophysiognomy</th>
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<th>P</th>
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</table>

1MA and MU: Mata Alta and Mussununga, respectively. 2Mean obtained from total litterfall of several species. 3Mean obtained from different fractions (leaves, twigs, reproductive organs and fragments) of fine litterfall. 4Site 1 and Site 2: most advanced and least advanced secondary successional stage, respectively.
phytophysiognomy (21.9 Mg ha$^{-1}$ vs. 6.1 Mg ha$^{-1}$). The high litter accumulation on the soil surface seen in the Mussununga area (which depends on the relationship between litterfall and its decomposition) reflects its slow decomposition there, in contrast to the results of the Mata Alta area, which showed higher rates of both litterfall and decomposition. As a result, the higher litter accumulation in the Mussununga area serves as a reservoir of mineral nutrients overlying a nutrient-poor sandy soil (therefore similar to restinga soils).

There is an equilibrium between the low litterfall rate and slow decomposition in restinga areas that minimizes nutrient losses through the gradual release of nutrients from the organic material (Camara et al., 2018a). The slow decomposition reflects a litter poor in nutrients, as seen in the present study, but rich in lignin, with elevated lignin:nitrogen and/or carbon:nitrogen ratios (Bonanomi et al., 2013). The leaves demonstrate higher lignification as a mechanism to minimize water losses through evapotranspiration, in restinga (Pereira et al., 2012).

Garay et al. (1995) found that approximately 30% of the fine roots of the plants growing in the Mussununga area were present in the litter standing stock, however this was not observed in the Mata Alta. The presence of fine roots (78-94% of total) in the topsoil (0.00-0.05 cm) in tropical forests (Rosado et al., 2011) represents an important mechanism linked to the optimization of nutrient absorption through direct absorption from decomposing litter. This ecological mechanism minimizes nutrient losses through leaching, especially in nutrient-poor sandy soils, thus favoring their retention in the ecosystem.

5. CONCLUSIONS

Fine aboveground litterfall, with a predominant leaf fraction, was continuous throughout the study years. However, higher litter deposition occurred in the rainy season in both phytophysiognomies.

Lower values for fine litterfall, nutrient input, and leaf litter decomposition were observed in the Mussununga area, probably functioning as nutrient conservation strategies, and thereby contributing to the ecological functioning of this ecosystem where soil fertility was lower than in the Mata Alta. Thus, the results corroborated the hypothesis that such aspects of nutrient cycling differ between these two contiguous physiognomies of Atlantic Forest with different floristic and edaphic conditions.

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